

## Article

# Sediment Transport Model Including Short-Lived Radioisotopes: Model Description and Idealized Test Cases

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Received: 1 October 2018; Accepted: 24 November 2018; Published: 27 November 2018



**Abstract:** Geochronologies derived from sediment cores in coastal locations are often used to infer event bed characteristics such as deposit thicknesses and accumulation rates. Such studies commonly use naturally occurring, short-lived radioisotopes, such as Beryllium-7 ( $^7\text{Be}$ ) and Thorium-234 ( $^{234}\text{Th}$ ), to study depositional and post-depositional processes. These radioisotope activities, however, are not generally represented in sediment transport models that characterize coastal flood and storm deposition with grain size patterns and deposit thicknesses. We modified the Community Sediment Transport Modeling System (CSTMS) to account for reactive tracers and used this capability to represent the behavior of these short-lived radioisotopes on the sediment bed. This paper describes the model and presents results from a set of idealized, one-dimensional (vertical) test cases. The model configuration represented fluvial deposition followed by periods of episodic storm resuspension. Sensitivity tests explored the influence on seabed radioisotope profiles by the intensities of bioturbation and wave resuspension and the thickness of fluvial deposits. The intensity of bioturbation affected the persistence of fluvial event beds as evidenced by  $^7\text{Be}$ . Both resuspension and bioturbation increased the modeled seabed inventory of  $^{234}\text{Th}$ . A thick fluvial deposit increased the seabed inventory of  $^7\text{Be}$  and  $^{234}\text{Th}$  but mixing over time greatly reduced the difference in inventory of  $^{234}\text{Th}$  in fluvial deposits of different thicknesses.

**Keywords:** numerical model; sediment transport; marine; short-lived radioisotopes

## 1. Introduction

Radioisotopic tracers, such as Beryllium-7 ( $^7\text{Be}$ ) and Thorium-234 ( $^{234}\text{Th}$ ) have been used to characterize sediment provenance, transport pathways, deposition, and biological mixing for numerous marine environments [1], including estuaries (e.g., [2–4]) and continental shelves adjacent to rivers (e.g., [5–7]). Both  $^7\text{Be}$  and  $^{234}\text{Th}$  are highly particle-reactive and therefore useful as tracers of sediment processes [8,9]. In river-dominated sedimentary environments, precipitation and subsequent river runoff can supply  $^7\text{Be}$ -laden sediments to coastal waters, so researchers have inferred that seabed sediment with significant  $^7\text{Be}$  activities contained recent terrestrial material [5,6,9–11]. Additionally, the decay profile of  $^7\text{Be}$  has been used to infer seabed mixing coefficients [9,12]. Th-234 is continuously produced in the coastal ocean via decay of its parent Uranium-238 [6,13,14]. It has been used both as a tracer of biological mixing intensity [15,16] and an indicator of recent deposition of (re-)suspended material [6,13]. Given the short half-lives of  $^7\text{Be}$  and  $^{234}\text{Th}$ , 53.3 and 24.1 days,

respectively, these radioisotopes provide useful indicators of processes that occurred over the last 100–250 days [9,11,13,14].

Several factors influence sediment bed profiles of radioisotopic tracers, including source terms; physical and biological mixing within the seabed; and time-scales of deposition, erosion, burial, and radioisotope decay [1,5,17]. For example,  $^7\text{Be}$  deposition from the atmosphere depends on seasonality of both precipitation and atmospheric inventory [18]. Scavenging of  $^{234}\text{Th}$  from the water column is more likely under high suspended sediment concentrations, such as during elevated river discharges or high wave energy, which both vary seasonally and with weather [6,19]. Sediment bed surface activities of  $^7\text{Be}$  in continental shelf environments range from 1 to 10 disintegrations per minute per gram of sediment ( $\text{dpm g}^{-1}$ ) and for  $^{234}\text{Th}$  range from 10 to 80  $\text{dpm g}^{-1}$  [5,6,19,20]. Specific activities of suspended sediments range from 0 to 50  $\text{dpm g}^{-1}$  for  $^7\text{Be}$  and 15 to 115  $\text{dpm g}^{-1}$  for  $^{234}\text{Th}$  under near oceanic salinities [2,3].

Bioturbation modifies the physical, chemical, and biological properties of sediment and can alter the radioisotope signal within surficial sediment deposits [12,16,17,21]. Upper continental shelf bioturbation rates are  $10\text{--}100\text{ cm}^2\text{ yr}^{-1}$  depending on water depth, seasonal conditions, location, and the specific radioisotope tracer used to infer them [12,17,22]. Many studies have estimated deposition and flood sediment budgets based on analysis of radioisotopic data (e.g., [6,7,9,23,24]). Additionally, deposition and mixing of seafloor sediment frequently occur at the same time, complicating the interpretation of radionuclide activity profiles within the seabed [17,23–26]. Although radioisotopes are useful tools for quantifying rates of biological mixing and short-term deposition, the complex combinations of processes can yield ambiguity in the interpretation of activity-depth profiles.

Numerical models have long promised to help improve our interpretations of radioisotopic profiles, but few models have directly represented the underlying physical and biological processes that shape activity-depth profiles. A previous study used a one-dimensional sedimentation–bioturbation model to estimate the fraction of the original unit volume of the seabed that retained its primary depositional fabric, defining this “the preservation quotient” [27]. The model included changes to the preservation quotient via depositional events and bioturbation processes, but neglected physical mixing, erosion, consolidation, and temporal variability [27]. Other researchers used a two-dimensional numerical model to represent gravity-driven sediment transport and deposition on the Waiapu shelf, New Zealand (NZ) [28]. The fluvial deposit thickness was multiplied by a decay factor to characterize a relative  $^7\text{Be}$  activity within the simulated flood deposit, which was then compared to available  $^7\text{Be}$  observations on the shelf [28,29]. Neither of these examples, however, directly modeled the seabed behavior of the radioisotopes.

More typically, three-dimensional sediment transport models simply calculate erosion and deposition, changes to sediment deposit thickness, and grain size distributions (e.g., [30–32]). Such calculations, however, are difficult to directly relate to radioisotopic records, even for cases where the models resolve both the temporal scales (i.e., event to seasonal timescales), and spatial scales (i.e., ~centimeter-scale sediment bed layers) at which the radioisotopic tracers are observed. For example, radioisotopic evidence of flood deposition for the Waipaoa shelf, NZ [7] was compared to model estimates for the same flood, but the modeled deposits were thinner than those inferred from the penetration depth of  $^7\text{Be}$  [33]. The mismatch was attributed to uncertainties related to sampling resolution, spatial patchiness in deposition, and physical and biological mixing of the radioisotope in the seabed [33]. Similarly,  $^7\text{Be}$  was used to identify flood deposit geometry and mass from Tropical Storm Lee in the Chesapeake Bay, and those observations compared favorably with model simulations of sediment dynamics for the event [34]. These modeled estimates of sediment reworking and deposition attempted to reproduce the observed scales of storm seabed-layer thickness subsequent to the peak storm erosion. However, there remains some ambiguity in the interpretation of observed seabed activity profiles because they also reflect downward mixing via physical reworking and biodiffusion that were unaccounted for in the model.

There is therefore a disconnect between the sediment transport models, which calculate sediment deposit thicknesses, and observations based on radioisotopic data which reflect deposition along with physical mixing and bioturbation. To our knowledge, no numerical sediment transport model has incorporated radioisotopic tracers, which would provide an obvious direct comparison to field observations. Additionally, few numerical sediment transport models include seabed mixing via bioturbation, which is necessary to directly compare modeled tracer profiles with seabed observations. Incorporating these processes into a numerical sediment transport model would help strengthen the ability to directly compare the modeled values and observations.

We modified a coupled sediment-transport and hydrodynamic model to account for the transport and decay of radioisotopes that were represented as reactive tracers in the water column and seabed. This paper describes the implementation of reactive tracers in the sediment-transport model and demonstrates the representation of short-lived radioisotopes in one-dimensional test cases of fluvial deposition and wave resuspension.

## 2. Materials and Methods

Radioisotopes were represented as reactive tracers in the water column and on the seabed through modifications to the Community Sediment Transport Modeling System (CSTMS) [32]. This section describes the model and the modifications that account for short-lived radioisotopes. It then outlines the one-dimensional (vertical), idealized model of fluvial deposition and wave resuspension used to demonstrate the model's representation of short-lived radioisotopes.

### 2.1. Adding Reactive Tracers to the CSTMS

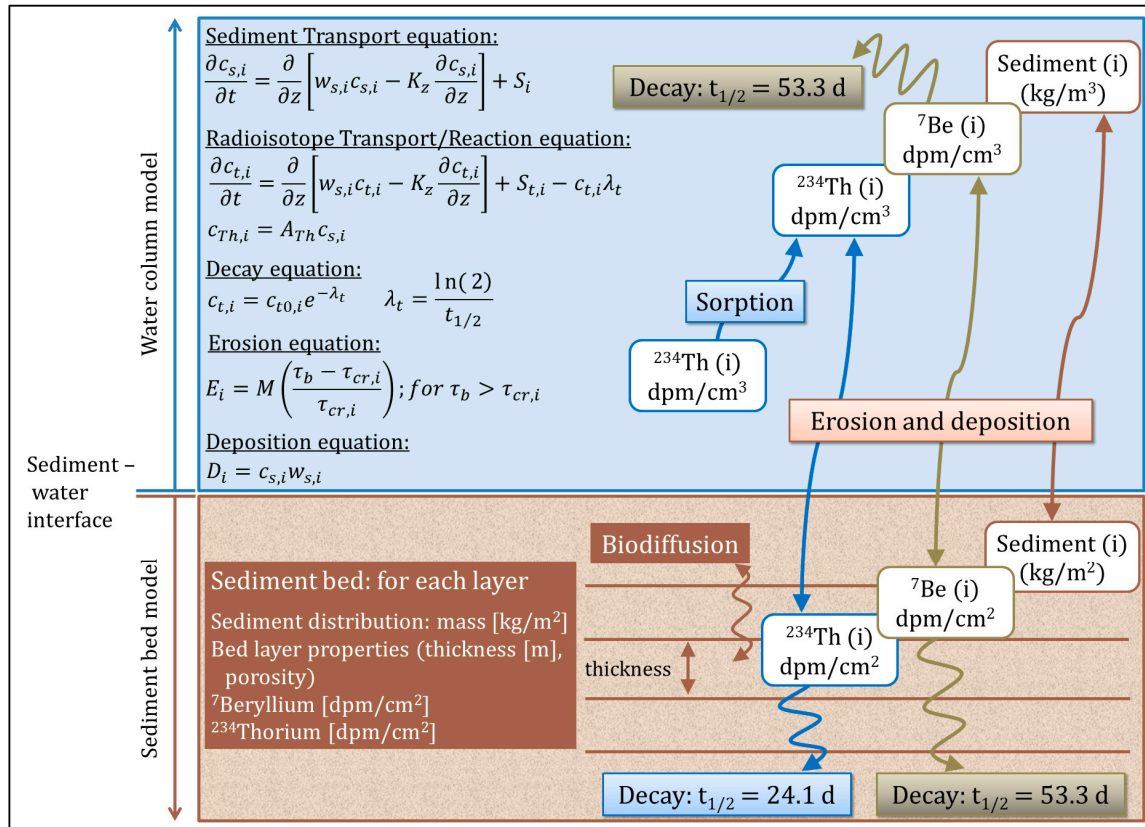
The CSTMS [32] operates as a module within the Regional Ocean Modeling System (ROMS; see [35,36]). Within the ocean modeling literature, a tracer refers to a dissolved or particulate scalar quantity that is transported by oceanic flows. Within ROMS, tracers are used to calculate the time-dependent, spatially varying fields of temperature and salinity. ROMS also uses tracers to track concentrations of sediment classes when the sediment module is used [32]; and biochemical constituents (oxygen, phytoplankton, etc.) when the biogeochemical module is used (e.g., [37]). These concentration fields are updated by solving mass balance equations that may include reactive terms for biogeochemical tracers [35,36]. Until recently, however, all implementations of the CSTMS assumed that the sediment classes acted as conservative tracers, so that their water-column concentrations were updated using the transport equation, with erosion and deposition terms for sources and sinks, and no reaction terms.

The sediment model uses tracers to account for a user-specified number of sediment classes (typically size classes) and represents grain size distributions on the seafloor using a user-specified number of sediment bed layers [32,38]. As a model run progresses, bed layer thicknesses and grain size distributions are adjusted to account for erosion and deposition. Additionally, changes to the thickness of the active transport layer, the layer of sediment at the seabed surface available for erosion [39], can modify the surficial bed characteristics. The non-cohesive sediment transport model used here [32] assumes that the active transport layer usually thickens during times of high excess bed shear stress [39], however, the parameter governing the thickness of the active transport layer is poorly constrained. The model also represents bed armoring, which occurs when the more mobile sediment (lower critical shear stresses and settling velocities) has been eroded; leaving the less mobile sediment (higher critical shear stresses and settling velocities) on the bed to form a lag layer that shelters underlying fine sediment from being entrained into the water column (e.g., [40]).

CSTMS applications commonly treat sediment as a conservative tracer, where the only sources and sinks of sediment are at open boundaries and the sediment bed [32]. Sediment resuspension can be stated as a one-dimensional (vertical) mass balance equation:

$$\frac{\partial c_{s,i}}{\partial t} = \frac{\partial}{\partial z} \left( w_{s,i} c_{s,i} - K_z \frac{\partial c_{s,i}}{\partial z} \right) + S_i, \quad (1)$$

where  $z$  is the vertical coordinate for height above the seafloor (m, positive upwards);  $K_z$  is eddy viscosity ( $\text{m}^2 \text{s}^{-1}$ );  $w_{s,i}$  is settling velocity ( $\text{m s}^{-1}$ );  $c_{s,i}$  is the concentration ( $\text{kg m}^{-3}$ ); and  $i$  designates the sediment class (Figure 1).  $S_i$  accounts for external sources or sinks of sediment. Typically,  $S_i$  includes fluxes into and out of the bottom cell to account for erosion and deposition. Note that the actual mass balance equation used in ROMS is three-dimensional and discretized for the model coordinate system.



**Figure 1.** Schematics of the sediment transport and geochronology model. Water column layers (blue) overlie seabed layers (brown) of variable thickness. Shaded boxes indicate processes; white boxes indicate state variables (tracers).

The sediment transport model has recently been enhanced to include reactive tracers whose transport equation includes reaction terms that represent geochemical processes that change (or exchange) quantities in tracer classes [38,41]. Reactive tracer quantities are transported in the water column at the same rate as associated water parcels or sediment with equivalent settling and erosion properties, so they can shadow a water mass or specific sediment class. The sediment bed model was also modified so that reactive tracers could be stored on the sediment bed and undergo exchanges between the water column and seabed. A recent application has used reaction terms to account for both particulate and dissolved geochemically reactive tracers in the seabed and the water column to evaluate the role of resuspension in biogeochemical cycles [42]. Here, we use a similar approach to simulate the transport, deposition, and decay of  $^7\text{Be}$  and  $^{234}\text{Th}$  (Figure 1), as described in the next section.

The one-dimensional (vertical) activity balance equation for a radioisotope is:

$$\frac{\partial c_{t,i}}{\partial t} = \frac{\partial}{\partial z} \left[ w_{s,i} c_{t,i} - K_z \frac{\partial c_{t,i}}{\partial z} \right] + S_{t,i} - c_{t,i} \lambda_t \quad (2)$$

where  $c_{t,i}$  is suspended tracer concentration ( $\text{dpm m}^{-3}$ ),  $w_{s,i}$  is the settling velocity ( $\text{m s}^{-1}$ ),  $K_z$  is eddy viscosity ( $\text{m}^2 \text{s}^{-1}$ ),  $S_{t,i}$  is the source/sink term of tracer concentration associated with the sediment class, and  $\lambda_t$  is the tracer decay constant ( $\text{s}^{-1}$ ). The index,  $i$ , denotes the sediment class to which this tracer is associated; while the index,  $t$ , denotes the reactive tracer. Aside from decay, tracers were assumed to remain associated with the sediment particles, due to the high particle-reactive nature of the tracers, so that desorption was neglected [1,8,9]. This behavior is consistent with previous studies in shelf environments (e.g., [8,43,44]). Because the transport terms ( $w_{s,i}$ ,  $K_z$ , and  $S_i$ ) act identically on the tracer class and the associated sediment class, this equation links the tracer activity to the mass concentration of the sediment class.

The boundary conditions for  $^7\text{Be}$  and  $^{234}\text{Th}$  were modified to represent their behavior in coastal environments. In our conceptual model,  $^7\text{Be}$  on river-dominated shelves is associated with fresh sediment delivered by fluvial input and is not significantly influenced by local atmospheric fallout [5]. We represented this by initializing the numerical model with sediment in suspension, notionally delivered by a fluvial source, and assigning this sediment a user-specified  $^7\text{Be}$  activity reflecting the fluvial provenance. As the model ran forward in time,  $^7\text{Be}$  activities behaved according to Equation (2) as  $^7\text{Be}$ -associated sediment settled, deposited, and possibly was resuspended.  $\text{Th-234}$  was treated differently. Our conceptual model was that  $^{234}\text{Th}$  is produced in seawater and scavenged by suspended particles [13,14]. We represented this in the numerical model by maintaining at a constant level the activity levels of  $^{234}\text{Th}$  associated with suspended sediment. In our model,  $c_{Th,i} = A_{Th}c_{s,i}$  (Figure 1); where  $A_{Th}$  was the assumed activity of suspended sediment ( $\text{dpm kg}^{-1}$ ) and  $c_{s,i}$  was the suspended sediment concentration ( $\text{kg m}^{-3}$ ). This approach was equivalent to assuming that the activity of any sediment in suspension immediately equilibrates to the level driven by local seawater activity. This assumes that equilibration in the model occurs over much shorter timescales than the timescales of typical resuspension events and natural removal of  $^{234}\text{Th}$  by sediment, which occurs on the order of days and is dependent on the amount of sediment resuspension [3,19,43]. Once sediment carrying  $^{234}\text{Th}$  enters the sediment bed, activities begin to decay. Through these treatments of the activities of  $^7\text{Be}$  and  $^{234}\text{Th}$ , we can characterize their behavior in a coastal ocean environment with river input and episodic resuspension.

The model tracked sediment mass and reactive tracer activities in terms of concentration ( $\text{kg m}^{-3}$  and  $\text{dpm m}^{-3}$ ) in each grid cell of the water column, and as layer-integrated quantities ( $\text{km m}^{-2}$  and  $\text{dpm m}^{-2}$ ) in each sediment bed layer. Erosion and deposition had corresponding influences on the inventory of radioisotope tracers and associated sediment in the water column and sediment bed. During erosion, reactive tracers linked to sediment classes were removed from the surface sediment layer (i.e., the active transport layer) and added to the water column. During deposition, reactive tracers linked to sediment were removed from the water column and returned to the surface sediment bed layer.

When the bed shear stress ( $\tau_b$ , Pascals, Pa) exceeds the critical shear stress ( $\tau_{cr,i}$ , Pa), the model estimates a sediment entrainment rate for each sediment class ( $i$ ) proportional to the excess shear stress ( $S = \tau_b - \tau_{cr,i}$ ) following the sediment entrainment equation [32]:

$$E_i = M \left( \frac{\tau_b - \tau_{cr,i}}{\tau_{cr,i}} \right), \quad (3)$$

where  $E_i$  is the entrainment rate ( $\text{kg m}^{-2} \text{s}^{-1}$ ),  $M$  is the erosion rate parameter ( $\text{kg m}^{-2} \text{s}^{-1}$ ), and  $i$  denotes the sediment class to which this tracer is associated. The model assumes simultaneous erosion and deposition (Figure 1), with the deposition equal to the settling flux [32]:

$$D_i = c_{s,i}w_{s,i}, \quad (4)$$

where  $D_i$  is the deposition rate ( $\text{kg m}^{-2} \text{s}^{-1}$ ) and  $c_{s,i}$  is the suspended-sediment concentration ( $\text{kg m}^{-3}$ ) in the bottommost grid cell. Like most implementations of the CSTMS, we assumed that the



hydrodynamic properties of sediment (i.e., settling velocity  $w_{s,i}$  and critical shear stress for motion  $\tau_{cr,i}$ ) remain fixed for each sediment class [32]. This neglects aggregation and disaggregation; settling velocities are fixed for each class because there is no exchange between sediment classes; though these have recently been included in other versions of the CSTMS [38].

We also modified the CSTMS sediment bed model to better represent stratigraphic processes that affect the distribution of radioisotopes in the seabed. The default sediment bed-layering scheme created new bed layers by dividing the bottom layer in half [32]. This can lead to having thick layers near the bed surface, and exponentially thinner layers with depth, so as in our recent work, we modified the scheme to form new layers by peeling off a thin layer from the bottom layer [38,41]. Additionally, we added biodiffusive mixing between bed layers of tracer concentrations and sediment types using a depth-dependent biodiffusion coefficient (see [38]). The biodiffusion coefficient,  $D_b$ , generally decreased with depth in the bed until it reached a small value specified in the input files.

Numerical solution of the tracer conservation equations was verified by comparison with the analytical equation for pure decay of initial activity (no erosion, deposition, or mixing) [45,46]. Using typical values for radioisotope activities and decay constants, burial rates, burial depths, and time scales; the model results were found to match analytical solutions [45,46] within 0.1 percent.

## 2.2. Idealized Test Case: Model Implementation

This paper presents simulations that explore behavior of the radioisotope tracers within a one-dimensional (vertical) model that represents the water column and underlying sediment bed (Figure 1). Configured similarly to a previously published model [38], our implementation included deposition of a thin fluvial layer and subsequent reworking via idealized storm forcing. Compared to the Standard Model Case, other runs varied parameters to explore sensitivity to the biodiffusion coefficient, resuspension intensity (maximum bed shear stress), and flood deposit thickness. For full details of the model implementation, model code, and input and output files, see the archive [47].

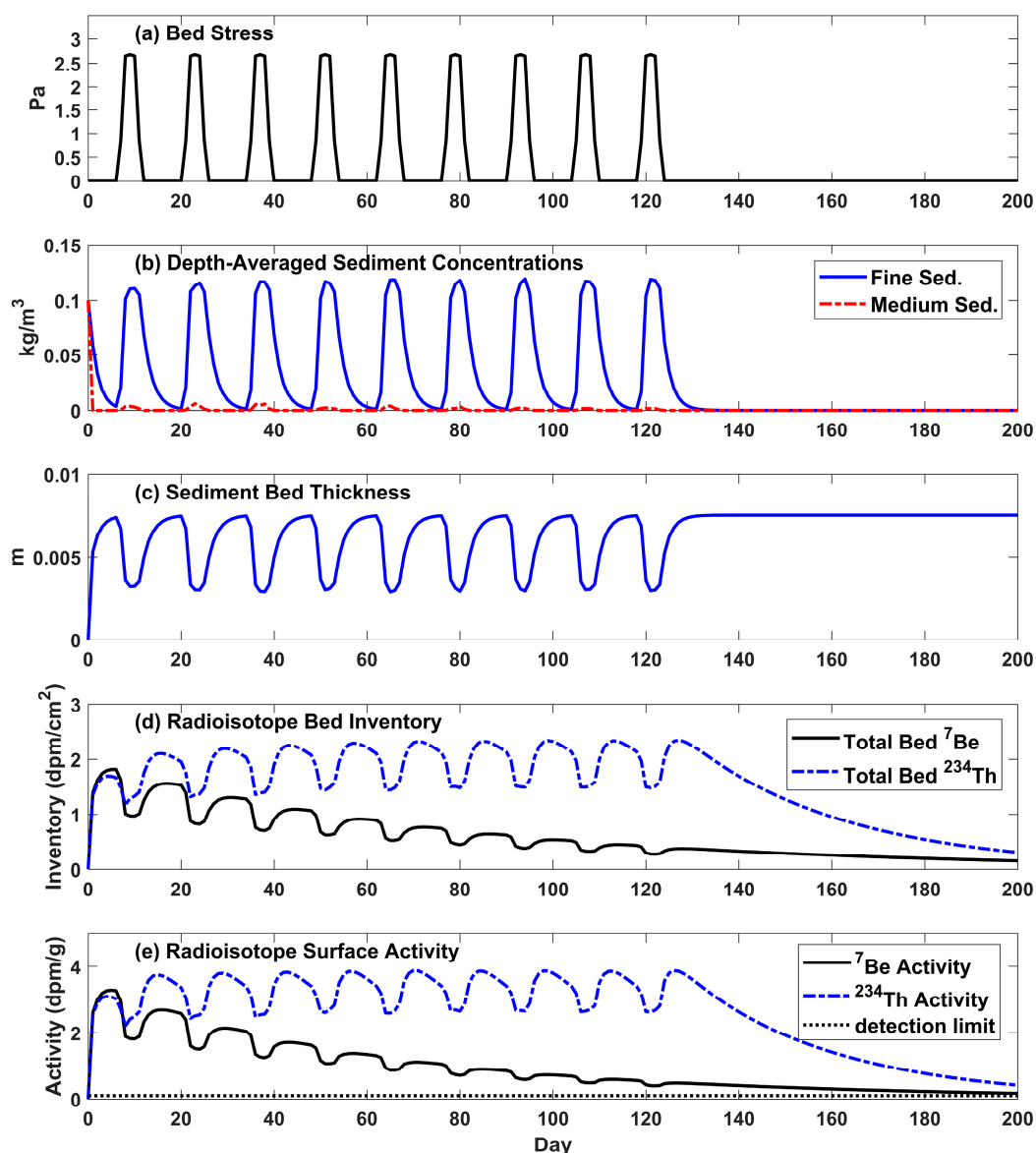
Specifically, the one-dimensional model represented a 20-m deep site using 30 water column layers. Several parameters were based on those used within a three-dimensional ROMS model for the northern Gulf of Mexico [48], including seabed porosity (0.8), and erosion rate parameter ( $M$ ;  $1 \times 10^{-5} \text{ kg m}^{-2} \text{ s}^{-1}$ ). The sediment model included three classes: fine (micro-floc), medium (macro-floc), and coarse (sand) with equivalent grain diameters of 15, 63, and 125 micrometers ( $\mu\text{m}$ ), respectively. These sediment classes assumed a quartz density ( $2650 \text{ kg m}^{-3}$ ) and had critical shear stresses set to 0.03, 0.08, and 0.1 Pa; and settling velocities of 0.01, 0.1, and  $1.0 \text{ cm s}^{-1}$ , respectively. The two fine classes represented micro-flocs and macro-flocs, though this version of the model neglected flocculation processes; i.e., the hydrodynamic properties of each class remained fixed and mass was not exchanged between classes. Radioisotopes  $^7\text{Be}$  and  $^{234}\text{Th}$  were linked to the two fine classes and assigned decay constants of  $\lambda_{Be} = 1.51 \times 10^{-7} \text{ s}^{-1}$  and  $\lambda_{Th} = 3.33 \times 10^{-7} \text{ s}^{-1}$ . The radionuclide adsorption properties were assumed identical for the fine and medium classes.

The upper 10 cm of the sediment bed were represented using 40 sediment layers, most of which were  $\sim 5 \text{ mm}$  thick. The initial sediment bed grain size distribution was assumed uniform with depth. Unless otherwise specified, biodiffusion occurred throughout the model runs. The biodiffusion coefficient was assumed to vary with depth [38]; for the Standard Case it was set to be  $1 \text{ cm}^2 \text{ yr}^{-1}$  from the sediment–water interface to a depth of 3 cm, then decreased linearly between 3 to 6 cm below the sediment–water interface. Below 6 cm, it was held constant at  $0.02 \text{ cm}^2 \text{ yr}^{-1}$ .

The water column initially carried suspended concentrations of  $0.1 \text{ kg m}^{-3}$  for each of the two finest sediment classes, which were tagged with radioisotopes having specific activities of both  $^7\text{Be}$  and  $^{234}\text{Th}$  of  $5 \text{ dpm g}^{-1}$ , a reasonable value for continental shelves (see Introduction). This material settled out of the water column within the first few days of the model, forming a flood deposit. The initial sediment bed had zero radioisotope inventories. The idealized model was configured to represent a situation where a single flood deposit is reworked by subsequent storm waves. Therefore,  $^7\text{Be}$  was added only at the beginning of the model, and the  $^7\text{Be}$  inventory subsequently decayed. In contrast,

the model assumed that  $^{234}\text{Th}$  activities associated with seabed sediment underwent decay, but that  $^{234}\text{Th}$  was replenished upon sediment resuspension. To represent this, the activity of  $^{234}\text{Th}$  on all suspended sediment was instantly reset to  $5 \text{ dpm g}^{-1}$ .

After the initial suspended sediment settled, the freshly-deposited fluvial layer was subjected to a 125-day idealized stormy period comprised of consecutive two-week intervals that each included a three-day storm and an eleven-day quiescent period (Figure 2a). Sediment was eroded under high bed shear stresses during the storms (2.7 Pa), and then settled and redeposited after bed stresses returned to background levels. The model neglected horizontal flux convergence or divergence, and therefore all sediment eroded during a storm redeposited when conditions subsided. The end of the model included a long quiescent period to evaluate model behavior (Figure 2a–c). Though idealized, the time-series for the wave forcing allows us to examine model behavior for a relatively simple case and has proved useful in past modeling studies [38].



**Figure 2.** Time series for the Standard Case (Case 1) of (a) bed stress, (b) depth-averaged suspended sediment concentrations (see legend), (c) bed thickness relative to initial, (d) radioisotope bed inventories, and (e) radioisotope surface activities. Shows first 200 days of 365-day run. Conditions remained quiescent after Day 130.

Three sets of experiments used the idealized scenario described above to investigate the sensitivity of seabed radioisotope activities to biodiffusion, flood layer thickness, and resuspension intensity. The Standard Case (Case 1) used intermediate values; while Cases 2–7 each varied one parameter (see Table 1). The first set of experiments tested a range of values for biodiffusion coefficients ( $D_{b,max}$ ) equal to 0, 1, and  $25 \text{ cm}^2 \text{ yr}^{-1}$ ; for Cases 2, 1, and 3, respectively. The second set varied flood input and considered a very thin (0.38 cm, Case 6), thin (0.75 cm, Case 1), and thicker (1.5 cm, Case 7) initial fluvial deposit. The influence of resuspension intensity was considered, both by varying the bed shear stress to include no (0 Pa, Case 4), moderate (2.7 Pa, Case 1), and intense (6.0 Pa, Case 5) resuspension; and by evaluating the sensitivity of radioisotopic profiles to the specification of the active transport layer thickness, i.e., the depth from which sediment can be recruited for resuspension during a model timestep. To prevent the active transport layer from being overly large for Cases 1–7, it was limited to a maximum of 0.5 cm below the sediment bed surface. Case 8 did not limit the active transport layer thickness, but implemented the commonly-used stress-dependent formulation (see [32,39]), via which the active layer thickness reached 1.8 cm.

To analyze the radioisotope behavior, model output was characterized in terms of the calculated penetration depth, radioisotopic inventory, detection timescale, and surface activity; we define those terms here. Surface activity ( $\text{dpm g}^{-1}$ ) represented the average activity in the top 1 cm of the seabed. The radioisotopic inventories ( $\text{dpm cm}^{-2}$ ) represented the depth integrals in the bed per cross-sectional area. Penetration depths (cm) of  $^7\text{Be}$  and  $^{234}\text{Th}$  have been used to infer the thicknesses of flood and storm deposits, and for each model run were characterized by the maximum depth below the sediment surface for which the  $^7\text{Be}$  and  $^{234}\text{Th}$  activities exceeded an assumed detection limit of  $0.1 \text{ dpm g}^{-1}$ . To explore how biodiffusion, resuspension intensity, and flood deposit thickness influenced the likelihood of detection using  $^7\text{Be}$  or  $^{234}\text{Th}$ , we evaluated the detection timescale, calculated as the length of time for which surface activity exceeded  $0.1 \text{ dpm g}^{-1}$ . The detection timescales of fluvial deposits were calculated as time elapsed since the flood event (i.e., in this case, the first day of the model run) based on detectability of  $^7\text{Be}$ . For storm reworking, the detection timescale of  $^{234}\text{Th}$  was calculated relative to the end of the last resuspension event (Day 125, except for the no resuspension Case 4). We define the thickness of the physically reworked material (cm) as the sum of the thicknesses of the resuspended material and the active transport layer; because this is the thickness of the sediment bed impacted during a storm event by erosion, redeposition, and armoring.

**Table 1.** Parameters for each model run. Bold print marks the parameters that differ for Cases 2–8.

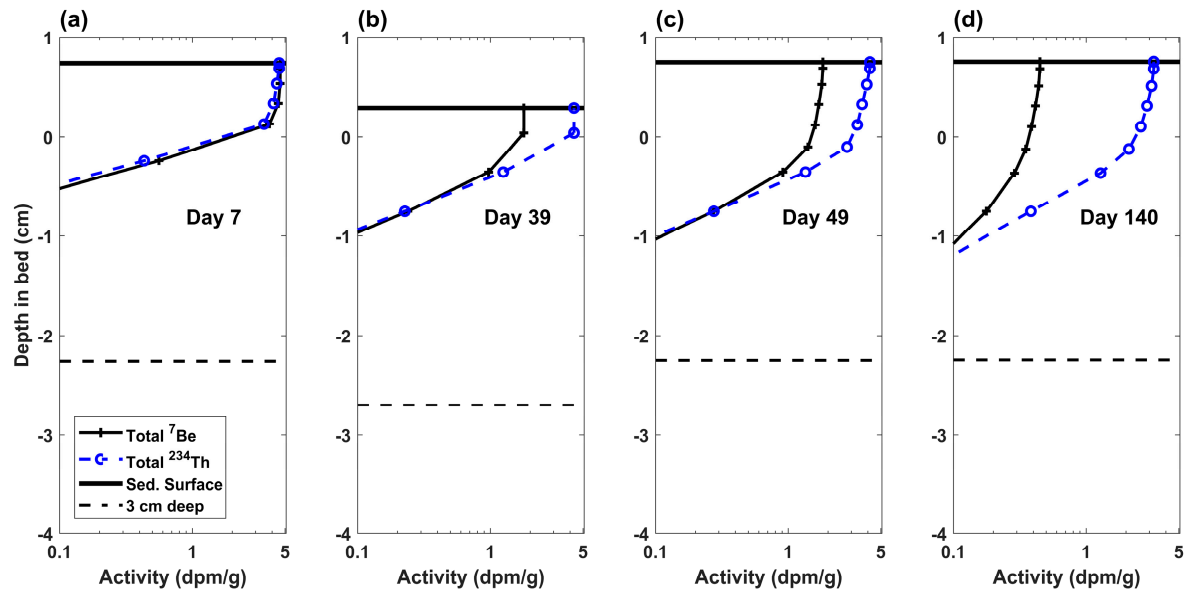
Case	$D_{b,max}$ ( $\text{cm}^2 \text{ yr}^{-1}$ )	Bed Stress (Pa)	Flood Layer (cm)	Active Layer (cm)
1	1	2.7	0.75	0.5
2	<b>0</b>	2.7	0.75	0.5
3	<b>25</b>	2.7	0.75	0.5
4	1	<b>0</b>	0.75	0.5
5	1	<b>6</b>	0.75	0.5
6	1	2.7	<b>0.38</b>	0.5
7	1	2.7	<b>1.51</b>	0.5
8	1	2.7	0.75	<b>1.8</b>

### 3. Results

The  $^7\text{Be}$  and  $^{234}\text{Th}$  activity profiles changed in response to episodic deposition and resuspension, and persistent biodiffusion and decay (Figure 3). The initial, 0.75 cm-thick, fluvial deposit contained uniformly high  $^7\text{Be}$  and  $^{234}\text{Th}$  activity (Figure 3a). During erosional episodes, activity profiles in the bed changed as sediment that carried radioisotopes was resuspended (Figures 2 and 3). Upon redeposition, profiles for both radioisotopes showed a surface layer that had uniform activities overlaying sediment whose activities decayed with depth (Figure 3c). Biodiffusion mixed both radioisotopes into the bed throughout the run, which acted to reduce radioisotope activities in the surface layer. As expected, bed inventories of  $^7\text{Be}$  and  $^{234}\text{Th}$  become different during episodes of resuspension, as  $^{234}\text{Th}$  was



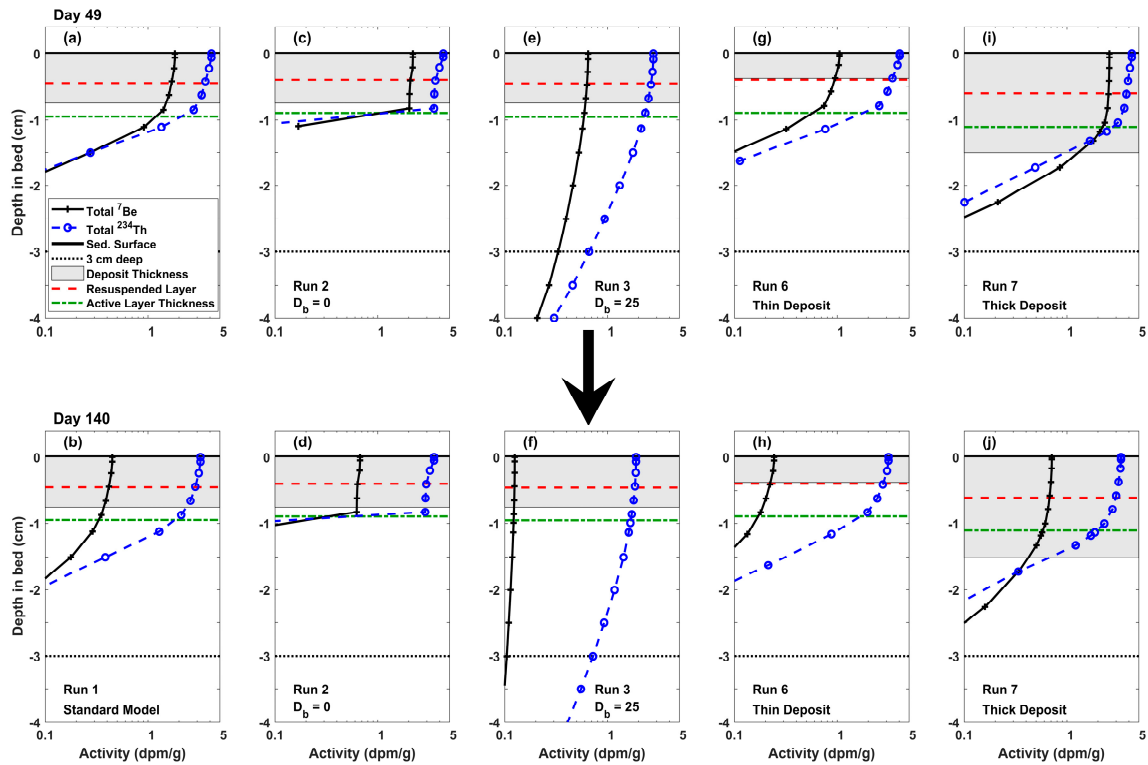
adsorbed from seawater onto the fine and medium sediment with each storm (Figure 3c,d). In contrast, storm resuspension had no net effect on  $^7\text{Be}$  bed inventory so that, over time,  $^7\text{Be}$  inventories decreased via decay, and surface activities were diluted via biodiffusion. Late in the model, after Day 130, inventories decreased due to the decay of  $^7\text{Be}$  and  $^{234}\text{Th}$ , and surface tracer activities were reduced via biodiffusion (Figure 2d,e).



**Figure 3.** Vertical profiles of seabed radioisotope activities at four instances for Case 1. (a) Day 7 after initial deposition, (b) Day 39 at peak of a resuspension event, (c) Day 49 after redeposition, and (d) Day 140 during the extended quiescent period. Solid black lines represent bed surface and thick dashed lines represent three centimeters deep in the bed. The bed height prior to initial deposition was defined as  $z_{bed} = 0$ .

### 3.1. Sensitivity to Biodiffusivity

Results from model runs that used different biodiffusion coefficients (Cases 1, 2, and 3; Table 1) illustrate the influence of biodiffusion intensity on radioisotope bed profiles. Figure 4a–f shows  $^7\text{Be}$  and  $^{234}\text{Th}$  profiles for these runs after the third and seventh resuspension events (Days 49 and 140, respectively). With increasing biodiffusion intensity, surficial sediments that carry high radioisotope activities were mixed into the seabed at faster rates, increasing the penetration depths for both  $^{234}\text{Th}$  and  $^7\text{Be}$ , and correspondingly decreasing surface activities. However,  $^{234}\text{Th}$  surface activities decreased less than  $^7\text{Be}$  because resuspended sediment adsorbed new  $^{234}\text{Th}$  from seawater, elevating bed inventories of  $^{234}\text{Th}$  as biodiffusion intensity increased. The  $^{234}\text{Th}$  and  $^7\text{Be}$  detection timescales of flood and storm beds decreased with increasing intensity of biodiffusion (Table 2). As  $D_{b,max}$  increased, each radioisotope was mixed more rapidly into the bed, diluting surface activities by vertical mixing, so that surface activities fell below the detection limit more quickly.



**Figure 4.** Radioisotope profiles calculated that (a,b) used  $D_{b,max} = 1 \text{ cm}^2 \text{ yr}^{-1}$  and had a flood deposit 0.75 cm thick (Case 1), (c,d) neglected biodiffusion (Case 2), (e,f) used  $D_{b,max} = 25 \text{ cm}^2 \text{ yr}^{-1}$  (Case 3), (g,h) had a flood deposit 0.38 cm thick (Case 4), and (i,j) had a flood deposit 1.51 cm thick (Case 5). Top row shows results on Day 49 while bottom row shows Day 140. Horizontal lines are as described in Figure 3. The gray box represents the extent of the initial flood deposit, red line denotes depth of resuspended layer, green line denotes the maximum penetration of the active layer when sediment was resuspended. The bed height after initial deposition is defined as  $z_{bed} = 0$ .

**Table 2.** Surface activities, bed inventories, penetration depths, and detection timescales after deposition (detection limit =  $0.1 \text{ dpm g}^{-1}$ ) for  $^7\text{Be}$  and  $^{234}\text{Th}$  for all cases. Days 49 and 140 represent times when no sediment was resuspended, but cover a range of time since initial deposition of  $^7\text{Be}$ .

	Day	Standard	Biodiffusion		Resuspension		Thickness		Active
		Case 1	Case 2	Case 3	Case 4	Case 5	Case 6	Case 7	Case 8
$^7\text{Be}$ Surf. Activity ( $\text{dpm g}^{-1}$ )	49	1.62	1.97	0.63	1.7	1.56	0.85	2.53	0.79
	140	0.4	0.61	0.12	0.43	0.39	0.21	0.68	0.22
$^7\text{Be}$ Bed Inventory ( $\text{dpm cm}^{-2}$ )	49	1.07	1.07	1.07	1.07	1.07	0.53	2.13	1.07
	140	0.33	0.33	0.33	0.33	0.33	0.16	0.66	0.33
$^7\text{Be}$ Penetration Depth (cm)	49	1.88	1.26	4.94	1.8	1.91	1.55	2.57	3.2
	140	1.88	1.07	3.44	1.77	1.91	1.39	2.56	2.97
$^7\text{Be}$ TD (months)		8.0	9.6	5.2	8.2	8	6.6	9.4	6.8
$^{234}\text{Th}$ Surf. Activity ( $\text{dpm g}^{-1}$ )	49	3.41	3.51	2.59	0.82	3.56	3.04	3.85	3.7
	140	2.72	2.87	1.82	0.05	2.86	2.52	3	2.96
$^{234}\text{Th}$ Bed Inventory ( $\text{dpm cm}^{-2}$ )	49	2.07	1.87	3.19	0.51	2.27	1.8	2.75	4.58
	140	1.75	1.52	2.74	0.04	1.95	1.58	2.1	3.73
$^{234}\text{Th}$ Penetration Depth (cm)	49	1.86	1.09	5.03	1.48	1.93	1.69	2.25	3.42
	140	1.97	1.09	5.7	0	2.11	1.96	2.21	3.49
$^{234}\text{Th}$ TD (months)		4.2	4.6	3.2	4.1	4.3	4.1	4.3	4.4

### 3.2. Sensitivity to Fluvial Deposit Thickness

Cases 6 and 7 had initial fluvial deposit thicknesses of 0.38 and 1.5 cm, one-half and double the deposit thickness of Case 1, respectively. Because the fluvial deposit provided the initial source of radioisotopes to the seabed, the deposit thickness directly influenced penetration depth and bed inventory for both radioisotopes early in the model run (Figure 4a,g,i; Table 2). Physical mixing was especially important for Case 6, for which the thickness of the resuspended layer exceeded that of the initial fluvial deposit.

The  $^7\text{Be}$  surface activity remained higher over time for the thicker deposits than for the thinner deposit, but the differences did not scale linearly with deposit thickness due to physical and biological mixing (Figure 4; Table 2). Doubling the fluvial deposit thickness increased the  $^7\text{Be}$  surface activity by 56% and the  $^7\text{Be}$  penetration depth by 37% on day 49 of the model runs (Table 2). The initial fluvial thickness played an even smaller role in the detection timescale, with thicker deposits remaining detectable for only slightly longer than calculated for the thinner deposits; doubling the flood thickness increased the detection timescale for  $^7\text{Be}$  by roughly 18% (Table 2).

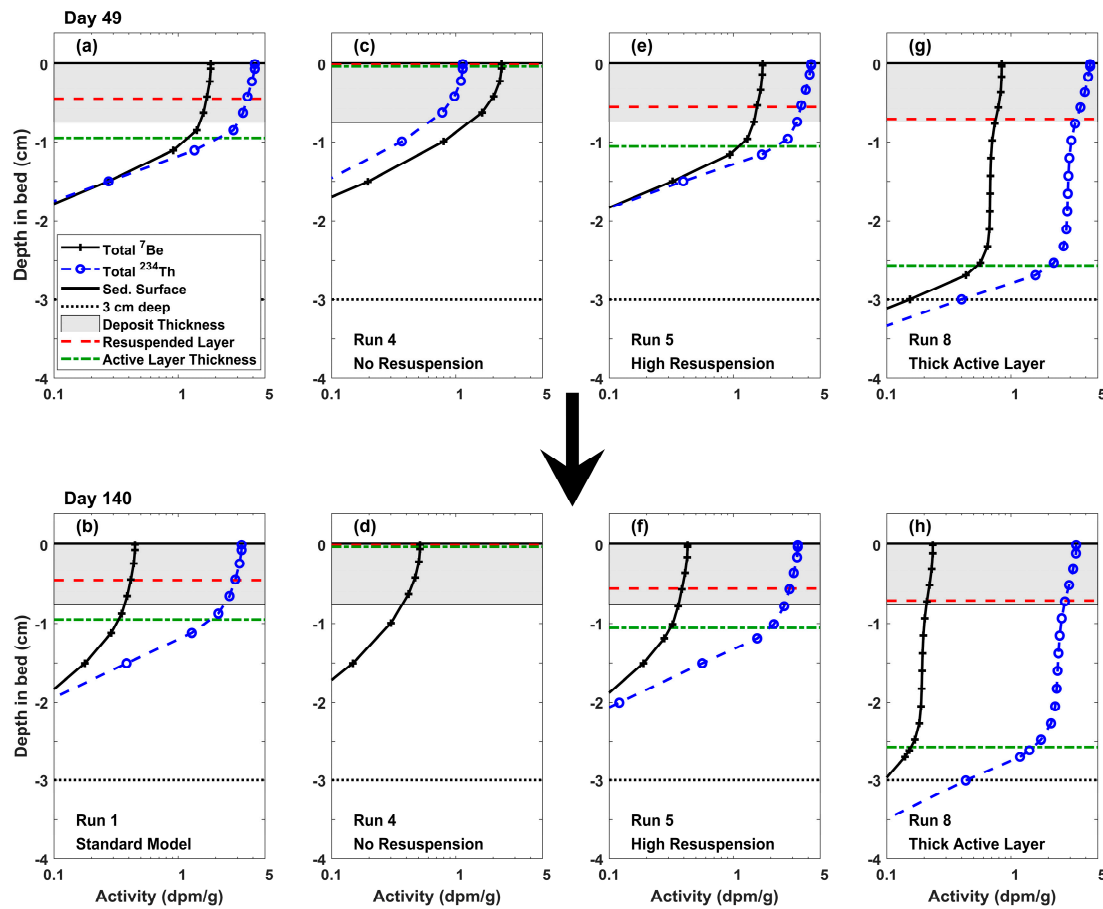
The relative thicknesses of the physically reworked layer compared with the initial flood deposit affected the response of the  $^{234}\text{Th}$  activities. Cases 1 and 6 had similar  $^{234}\text{Th}$  penetration depths (Figure 4a,b,g,h; Table 2) because they were set by the thickness of the physically reworked layers. The fact that the initial fluvial deposit thickness for Case 7 exceeded that of the resuspended layer directly affected the penetration depth (Figure 4a,i; Table 2). There was feedback between fluvial deposit thickness and mass of sediment resuspended, because the fluvial layer only contained fine and medium sediment. A thicker fluvial layer therefore supplied more fine and medium sediment near the surface, allowing more mobile sediment to be eroded and become tagged with fresh  $^{234}\text{Th}$  in the water column. This increased the  $^{234}\text{Th}$  inventories and penetration depths after resuspension events (Table 2).

### 3.3. Sensitivity to Resuspension Intensity

Cases 4 and 5 explored how the seabed profiles of  $^7\text{Be}$  and  $^{234}\text{Th}$  responded to changes in resuspension intensity by varying the magnitude of the maximum bed shear stress,  $\tau_b$ . The cases featuring zero (Case 4), medium (Case 1), and high resuspension intensity (Case 5) yielded erosion depths of 0, 0.46, and 0.56 cm, respectively (Figure 5). For the cases that included resuspension, the thickness of the fluvial deposit (0.75 cm) exceeded the peak erosion depths. Bed inventories of  $^{234}\text{Th}$  increased with resuspension intensity; energetic events suspended more sediment which then scavenged  $^{234}\text{Th}$  from the water column (Table 2). The case with no resuspension therefore had much lower  $^{234}\text{Th}$  surface activity and bed inventory than the other cases (Figure 5c,d). In contrast, resuspension did not affect  $^7\text{Be}$  bed inventories, as the model assumed that  $^7\text{Be}$  was only supplied in the initial flood deposit. Via physical mixing, erosion and redeposition cycles increased penetration depths for both radioisotopes. During periods of erosion, the active transport layers shifted deeper into the seabed, mixing radioisotope-tagged surficial sediments with underlying lower-activity sediment. This decreased surface activities and detection time-scales with increasing resuspension intensity for  $^7\text{Be}$ , because it was not replenished during storms.

Case 8 evaluated the role of the active layer thickness on resuspension intensity and its impact on modeled radioisotope profiles. While other cases limited the active transport layer thickness to a maximum of 0.5 cm, Case 8 used a widely accepted formulation (i.e., [39]) for it that reached 1.8 cm during the storm. Therefore, though bed stresses were identical, the amount of suspended sediment and the eroded depth was almost doubled in Case 8 compared to Case 1. The active transport layer thickness also greatly influenced the modeled radioisotope signatures (Figure 5a,b,g,h). While total  $^7\text{Be}$  bed inventory was insensitive to the active transport layer thickness, the increased erosion and concurrent mixing of the surficial few centimeters of the bed effectively diluted the surface activity for  $^7\text{Be}$  and increased the penetration depth (Figure 5a,b,g,h). In contrast, the surface activity of  $^{234}\text{Th}$  was insensitive to the increased active transport layer thickness, (Table 2, Figure 5a,b,g,h), but the resultant

higher resuspension depths nearly doubled the  $^{234}\text{Th}$  inventory and the penetration depths for Case 8 relative to Case 1 (Table 2, Figure 5a,b,g,h).



**Figure 5.** Radioisotope profiles that had (a,b) resuspension depth of 0.46 cm (Case 1), (c,d) no resuspension (Case 4), (e,f) resuspension depth of 0.56 cm (Case 6), and (g,h) a thick active layer (Case 8). Top row shows results on day 49; while bottom row shows day 140. Horizontal lines as described in Figure 3. Gray boxes represent the initial extents of the flood deposit, red lines denote depth of resuspended layer, green lines denote depth of maximum active layers. The bed height after deposition was defined as  $z_{\text{bed}} = 0$ . Note that day 140 for Run 4 (d) has zero  $^{234}\text{Th}$ , which decayed below the detection limit by this time.

## 4. Discussion

### 4.1. Synthesis of Sensitivity Tests

The results summarized in Table 2 demonstrate that variable bioturbation rates, resuspension intensity, initial flood deposit thickness, and active layer thickness affected modeled radioisotope profiles, inventories, penetration depths, and detection time scales to varying degrees. The bioturbation sensitivity tests (Cases 2 and 3) generated results that deviated most from Case 1. Note that these bioturbation rates spanned a larger range than considered in the other sensitivity trials, but this reflects the significant uncertainty in characterizing bioturbation coefficients (see [46]). In Case 3, enhanced bioturbation increased the  $^{234}\text{Th}$  and  $^{7}\text{Be}$  penetration depths, reduced surface activities, and caused the shortest detection timescales for any of the model runs (Table 2, Figures 4 and 5). Case 2 neglected bioturbation and produced the thinnest penetration depth and longest detection time scale of the cases considered (Table 2). Varying the initial fluvial layer thickness appreciably affected the  $^{7}\text{Be}$  results, especially  $^{7}\text{Be}$  bed inventory, more so than  $^{234}\text{Th}$ . Results from the cases that investigated resuspension intensity and active layer thickness suggested that increasing resuspension,

in the presence of bioturbation, worked to dissipate the terrestrial  $^7\text{Be}$  signal, while enhancing the oceanic  $^{234}\text{Th}$  signal.

The reactive tracer model responded reasonably to a range of dissipative and intensifying conditions for two radioisotopes that are indicative of terrestrial and oceanic sources. Analysis of the idealized test case results produced radioisotope profiles and metrics similar to those observed on continental shelf environments (e.g., [5,9,17]), and consistent with patterns noted in field observations. For example, bioturbative processes are known to mix radioisotopes into the seabed, increasing penetration depths and reducing detection time scales [16,17,23,25]. Similar to the results that varied fluvial layer thickness (Cases 6 and 7; Figure 4; Table 2), bed inventories of  $^{234}\text{Th}$  and  $^7\text{Be}$  have been observed to correlate positively with initial flood layer thickness [5]. Model experiments also provided scenarios wherein shorter-lived  $^{234}\text{Th}$  was observed at greater depth than the longer-lived  $^7\text{Be}$  (Cases 1, 3, 6, 8) as observed in the Adriatic Sea [9].

#### 4.2. Limitations of Model as Implemented Here

This model marks a first step in integrating sediment transport process models [32] with the radioisotopic tracers that have been increasingly used as indicators of sediment transport processes (e.g., [3,5–7,9,13]). This section reviews some of the assumptions that were made in developing this model and discusses their implications.

The model discussed herein used a relatively simple approach to represent behavior of radioisotopes, but integrated them into an established sediment transport model. In particular, the model employed a simple conceptual model for the sources of the radioisotopes, and for their interactions with sediment that would be appropriate for river-dominated areas where these radioisotopes have been used to infer source and transport processes. To represent an initial flood source for sediment, the only source of  $^7\text{Be}$  was assumed to be sediment suspended at the start of the model run, which had a prescribed activity of associated  $^7\text{Be}$ . While it is known that activities of fluvially discharged  $^7\text{Be}$  vary in time [9], our idealized scenario represented only an initial flood pulse so that temporal variation in  $^7\text{Be}$  was not an issue for the runs presented. Additionally, our conceptual model neglected other sources of  $^7\text{Be}$ ; for example, direct precipitation has been noted as a source of  $^7\text{Be}$  on non-river influenced continental shelves [49]. Similarly, the conceptual model simplified the processes that influence  $^{234}\text{Th}$  activities in coastal environments. The model also assumed that all suspended sediment carried a constant, pre-set activity of  $^{234}\text{Th}$ . This was equivalent to assuming that sorption of  $^{234}\text{Th}$  to sediment grains occurred instantaneously upon suspension in the model water column, and that there was no variation in background  $^{234}\text{Th}$  concentrations. In actuality, the association of  $^{234}\text{Th}$  with suspended sediment grains may take days to equilibrate, dependent on suspended sediment concentrations [3,19,43], and  $^{234}\text{Th}$  concentrations are known to vary with salinity within estuaries and on river-influenced shelves [1]. Additionally, the model assumed that the activities of the radioisotopes on the muddy sediment did not vary with grain size or water column properties.

The sediment transport model also included some simplifications. For example, the resuspension of muddy sediment is subjected to cohesive processes such as bed consolidation and flocculation. The version of the CSTMS used here [32] neglected these processes, so that the sediment classes had pre-set, constant values to represent erodibility and settling velocity. For application to reworking of muddy flood deposits, future studies might consider using the radioisotopic reactive tracer module presented here with the version of the CSTMS that also accounts for flocculation and bed consolidation [38].

Bedload was not directly represented in our model, though the CSTMS provides bedload formulations that can be applied to the upper-most sediment bed layer, i.e., the active transport layer [32]. We did not use the bedload transport formulations, however. Because our idealized one-dimensional (vertical) model assumes uniform horizontal sediment fluxes, there would be no net erosion or deposition from bedload. One conceptualization of the active transport layer, however, is as a bedload layer from which sediment can be suspended [40]. As bedload intensifies, the active layer thickens, which in our model acted to mix and dilute the radioisotope tracer concentrations in the



upper sediment bed. The sensitivity of surface activity and penetration depth to the formulation of the active transport layer (Cases 1 and 8) illustrates this.

#### 4.3. Future Applications

As discussed in the previous section, our idealized model used a relatively simple representation of radioisotopes and neglected some sediment transport processes. Though this model did not represent the full complexities of actual radioisotope source terms, or particle–radioisotope interactions, the model itself could provide a venue for future studies that consider the impact of these complexities on sediment bed geochronological tracer profiles. Additionally, by developing our coupled reactive tracer–sediment transport model within a community modeling framework, this paves the way for future studies to consider the potential impacts of other sediment transport processes or radioisotopic chemistry on the geochronological record.

In addition to idealized studies such as reported here, the geochronological model should provide useful insight if applied to realistic field settings. For example, we have used a similar one-dimensional (vertical) version of the model to represent a site offshore of the Mississippi subaqueous delta [50]. This model was implemented to replicate observed radioisotopic profiles [6]. The model-produced radioisotopic profiles were then used within various analytical models from the literature to derive deposition rate, etc., from the simulated profiles; which were then compared to the actual model deposition rates. Results showed that the analytical model estimates can be unreliable if assumptions regarding the biodiffusion rate and radioisotope activity of freshly deposited sediment are incorrect [50].

It would be useful to add the capability of our sediment transport geochronological model to a full three-dimensional model, such as that developed for the Waipaoa River Shelf, NZ [51], or a northern Gulf of Mexico continental shelf model [48,52]. Both of these efforts attempted to use radioisotopic data for comparison to model estimates of sediment deposition; but this exercise presented challenges that may be abated by direct inclusion of the radioisotopes within the transport equations. Accounting for geochronology within realistic domain models could facilitate model validation by enabling direct comparison between the modeled and measured quantities, i.e., profiles of  $^{234}\text{Th}$  and  $^7\text{Be}$ . Within a joint field and modeling study, this capability could enable careful model–data comparisons that could yield insight into the specification of some model parameters that have been difficult to constrain, such as the active transport layer thickness.

#### 5. Conclusions

A sediment transport model has been modified to account for reactive terms and represent transformations of particulate tracers in the water column and seabed. The model was applied to calculate sediment bed profiles of short half-life radioisotopes  $^7\text{Be}$  and  $^{234}\text{Th}$  that have been used to interpret deposits of river derived sediment and marine sediment resuspension, respectively. Examples were provided in an idealized one-dimensional (vertical) model that represented sediment suspension and the seabed, and that evaluated the response of radioisotope profiles to variations in riverine sediment input, storm intensity, and biodiffusive mixing coefficient. Radioisotopic values such as surface activity, inventory, and penetration depth were analyzed from model results and compared to the sediment input and reworking represented in the model. The intensity of biodiffusion affected the persistence of detectable fluvial event beds. Using a biodiffusion coefficient of  $25\text{ cm}^2\text{ yr}^{-1}$ , the fluvial event bed would be obscured within days or weeks. Both resuspension and biodiffusion increased the seabed inventory of  $^{234}\text{Th}$ . Biodiffusion diluted the surface layer by mixing sediment enriched with  $^{234}\text{Th}$  (and  $^7\text{Be}$ ) with underlying material. When surficial sediment underwent cycles of resuspension and redeposition, however, it became enriched in  $^{234}\text{Th}$  and thereby increased the total bed inventory upon redeposition.

**Author Contributions:** Conceptualization, J.J.B., C.K.H., C.R.S.; Methodology, J.J.B., C.K.H.; Software, J.J.B., C.K.H., C.R.S.; Validation, J.J.B., C.K.H., T.A.K.; Formal Analysis, J.J.B., T.A.K.; Investigation, J.J.B., T.A.K.;

Resources C.K.H.; Data Curation, J.J.B., C.K.H.; Writing-Original Draft Preparation, J.J.B., C.K.H.; Writing-Review & Editing, J.J.B., C.K.H., T.A.K., C.R.S.; Visualization, J.J.B.; Supervision, C.K.H.; Project Administration, C.K.H.; Funding Acquisition, C.K.H.

**Funding:** The Bureau of Ocean Energy Management (BOEM) provided funding for Birchler, Harris, and Kniskern. During his M.S. program Birchler received additional funds from VIMS' Office of Academic Studies. This work was partially supported by the U.S. Geological Survey, Coastal and Marine Geology Program.

**Acknowledgments:** Use of firm and product names is for descriptive purposes only and does not imply endorsement by the U.S. Government. The authors are thankful for input and support from G. Auad (BOEM). The authors appreciate input from VIMS' faculty members of Birchler's M.S. committee, namely C. Friedrichs, S. Kuehl, and L. Schaffner. The authors also appreciate input from an anonymous reviewer of an earlier version of this manuscript, two anonymous reviewers of this journal, and input from R.C. Mickey and J. Moriarty (USGS). We feel that their input helped to improve this submission.

**Conflicts of Interest:** The authors declare no conflict of interest.

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